Revisiting the conservation genetics of Pampas deer (Ozotoceros bezoarticus)

Susana González^{1, 2*}, Leticia Repetto¹, Verónica Gutiérrez¹, María Eugenia Olivera¹, Claudia Corbi Botto¹, Yanina Leone¹, Mariano L. Merino³, Fernanda Góss Braga,¹ José Mauricio Barbanti Duarte,⁴ Jesús E. Maldonado⁵, and Mariana Cosse¹

- ¹ Instituto de Investigaciones Biológicas Clemente Estable-Ministerio de Educación y Cultura. (IIBCE-MEC). Av. Italia 3318. 11600 Montevideo 11100, Uruguay. Email: <u>sgonzalez@iibce.edu.uy</u>(SG), <u>leticiarepetto@gmail.com</u> (LR), <u>verogcoppetti@gmail.com</u> <u>com</u>(VG), <u>maeugeniaolivera@gmail.com</u> (MEO), <u>claucorbi@gmail.com</u>(CCB), <u>yanileone11@gmail.com</u>(YL), <u>fgbbio@hotmail.com</u> (FGB), <u>mcosse@iibce.edu.uy</u> (MC).
- ²Deer Specialist Group, IUCN Species Survival Commission, Rue Mauverney 28, 1196. Gland, Switzerland. Email: <u>sgonzalez@iibce.edu.uy</u> (SG). ³Centro de Bioinvestigaciones, Universidad Nacional del Noroeste de la Provincia de Buenos Aires, Comisión de Investigaciones Científicas de la Provincia de Buenos Aires, Pergamino. Buenos Aires, Argentina. Email: <u>mariano.merino@nexo.unnoba.edu.ar</u> (MLM).
- ⁴Núcleo de Pesquisa e Conservação de Cervídeos (NUPECCE) Departamento de Zootecnia Faculdade de Ciências Agrárias e Veterinárias Universidade Estadual Paulista (UNESP) Via de acesso Paulo Donato Castellane, s/n CEP: 14884-900. Sao Paulo, Brazil. Email: <u>mauricio.barbanti@unesp.br</u> (JMBD).
- ⁵ Center for Conservation Genomics, Smithsonian National Zoo and Conservation Biology Institute. Washington D. C., U.S.A. Email: <u>maldonadoJ@si.edu</u>(JEM).

*Corresponding author: https://orcid.org/0000-0001-6470-6182.

The Pampas deer (*Ozotoceros bezoarticus*) is a unique species of neotropical cervid, that inhabits a wide range of open habitats including grasslands, pampas, savannas, and cerrado (Brazil) from -5° to -41° S. The reduction of the area encompassed by these habitats has been dramatically reduced to less than 2 % by human activities such as agriculture, urbanization, and poaching. Three decades ago, we initiated a molecular genetic study of the Pampas deer based on representative samples from throughout their geographic range. Our aim is to reevaluate the effect that habitat fragmentation has had on gene flow among eight wildlife Pampas deer populations and one from the captive breeding centre *Estación de Cría de Fauna Autóctona* (ECFA). We examined DNA sequences from three mitochondrial markers: the control region (*D-loop*), Cytochrome b (*Cytb*), and Cytochrome Oxidase I (*COI*). Furthermore, we compared the resolution of the different mitochondrial markers to elucidate the phylogenetic and phylogeographic patterns of the species to define Evolutionary Significant Units (ESU`s). The amount of gene flow was correlated with geographic distance among groups and populations and was consistent with limited dispersal being the primary determinant of genetic differentiation between populations. Our results showed that the *D-loop* was the most appropriate marker for defining Evolutionary Significant Units. We found that the critically endangered Paraná population showed significant genetic distance from the others and revealed unique haplotypes with all the mitochondrial markers. The molecular genetic results provide a mandate for habitat restoration and design a management plan to conserve these relictual populations.

El venado de las pampas (*Ozotoceros bezoarticus*) es la única especie de cérvido neotropical, que habita en una amplia gama de hábitats abiertos que incluyen pastizales, pampas, sabanas y cerrado (Brasil) desde -5° a -41° S. Se ha reducido drásticamente su hábitat a menos del 2 % por las actividades humanas como la agricultura, la urbanización y la caza furtiva. Comenzamos hace tres décadas un estudio de genética molecular del venado de las pampas basado en muestras representativas de todo su rango geográfico. Nuestro objetivo es el de reevaluar el efecto de la fragmentación del hábitat sobre el flujo de genes entre ocho poblaciones de ciervos de las pampas silvestres y una del centro de cría en cautiverio Estación de Cría de Fauna Autóctona (ECFA). Examinamos las secuencias de ADN con tres marcadores mitocondriales: la región de control (*D-loop*), el citocromo b (*Cytb*) y el citocromo oxidasa I (*COI*). Además, comparamos la resolución de los diferentes marcadores mitocondriales para dilucidar los patrones filogenéticos y filogeográfica entre grupos y poblaciones siendo consistente con la dispersión, la principal limitante y determinante de la diferenciación genética entre poblaciones. Nuestros resultados mostraron que *D-loop* es el marcador adecuado para definir Unidades Evolutivas Significativas. La población de Paraná se encuentra en peligro crítico de extinción al tener una distancia genética significativa de las demás y haplotipos únicos con los marcadores mitocondriales. Los resultados de la genética molecular proporcionan un mandato para la restauración del hábitat y el diseño de un plan de gestión para conservar estas poblaciones relictuales.

Keywords: Cervidae; COI; Cyt b; D-loop; genetic management units.

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Introduction

The Pampas deer (*Ozotoceros bezoarticus*) was once a widespread and abundant species occupying a wide range of open habitats, including grasslands, pampas in Argentina and the Brazilian savanna known as the Cerrado (<u>Cabrera</u> <u>1943</u>; Jackson 1987; Merino *et al.* 1997; <u>González *et al.* 2002</u>; <u>Weber and González 2003</u>; <u>González et al. 2010</u>). Historical records showed that this species had a wide range distribution in southeastern South America (from -5° to -41° S), reported by several naturalists mentioned in their records as Charles Darwin (<u>Darwin 1860</u>). The area encompassed by these habitats has been dramatically reduced to less

than 2 % by human activities such as agriculture, urbanization and poaching of that present in 1900 (González *et al.* 2023). Currently, Pampas deer populations are generally small and isolated (Jackson and Langguth 1987; Pinder 1994; González *et al.* 2010, Figure 1).

Mitochondrial DNA has been the most widely used tool for reconstructing population and species histories, presumably because it is relatively easy to amplify, typically non-recombining, supposedly nearly neutral, and highly variable between and within species (Taberlet 1996; Avise <u>1998</u>). Furthermore, this molecular marker of maternal inheritance is useful for the genetic analysis of populations in fragmented habitats, such as those of the Neotropical deer (Avise 1992, 1995). Most deer species usually have: 1) asymmetric genetic flow and dispersal rate, females frequently being phylopatric. 2) Females and fawns are spatial associated. 3) A strong maternal lineages structuration resulting in demographic autonomy among populations in an ecological scale (González et al. 1998; Márquez et al. 2006; González et al. 2010). Historical population sizes based on mtDNA control region sequences were estimated to be several magnitudes larger than present day estimates. Gene flow patterns also showed high levels of genetic differentiation among isolated populations using representative samples of Pampas deer from throughout its geographic range. In that study, González et al. (1998) identified five conservation genetic units for the six localities surveyed: two in Brazil (Emas National Park and Pantanal da Nhecolandia), two in Uruguay (Salto and Rocha), and one in Argentina (composed of two populations: Samborombón Bay and San Luis).

The comparative craniometric analysis of the different Pampas deer populations revealed that differentiation was concordant with the levels of genetic differentiation found with mitochondrial markers. Furthermore, two new subspecies were recognized in the Uruguayan northwestern (*O. b. arerunguaensis*, Salto Department) and eastern (*O. b. uruguayensis*, Rocha Department) grasslands (<u>González et al. 2002</u>).

This species is currently considered Near Threatened (NT) in the global IUCN Red List (González et al. 2016). In their southern range of South America, the most threatened populations occur in Argentina, Bolivia, Paraguay, and Uruguay with fewer than 2,500 mature individuals. This is reflected by the Red List categories of the subspecies: *O. b. celer*, -Argentina-: Endangered [EN B1ab(iii)]; *O. b. arerunguaensis* -Uruguay- [CR B1ab(iii); O. b. uruguayensis -Uruguay- [CR B1ab(iii)]; *O. b. bezoarticus* -Brazil- (DD); *O. b. leucogaster* - Argentina, Bolivia, Brazil; Paraguay- Near Threatened (NT) (González et al. 2016). The species is also included in CITES Appendix I (Giménez-Dixon 1987).

Three decades ago, we conducted the first molecular genetic study of the Pampas deer based on representative samples from throughout their geographic range (González et al. 1998). We aimed to deduce genetic units for conservation (Moritz 199514) and to better understand the effect of habitat fragmentation on gene flow and genetic variation. In this study, we evaluated the effects of habitat fragmenta-

tion on gene flow among eight wild Pampas deer populations and one from the captive breeding centre Estación de Cria de Fauna Autóctona (ECFA) in Uruguay. We examined DNA sequences from three mitochondrial markers: the control region (*D-loop*), Cytochrome b (*Cytb*), and Cytochrome Oxidase I (*COI*) to determine levels of genetic differentiation among isolated populations. Additionally, we compared the resolution of the different mitochondrial markers to elucidate the phylogenetic and phylogeographic patterns of the species and to define Evolutionary Significant Units (ESU`s).

Almost thirty years later, we increased the sample size from the previous locations and incorporated new sites from its wide geographic range to revisit the genetic characterization of Pampas deer by implementing additional mitochondrial markers. Our results will be providing a comprehensive approach for understanding the current genetic status and the future viability of the species.

Materials and methods

Sample collection. We analyzed 164 Pampas deer samples from eight geographic localities across the species range from Argentina, Bolivia, Brazil, and Uruguay and one captive population in Uruguay, ECFA (Figure 1; Supplementary Material Table 1). This captive stock was founded in 1980 with 10 individuals from the wild population from Salto.

DNA extraction and PCR amplification. Genomic DNA was extracted from tissue samples (50 mg, see details on Supplementary material) following González et al. (1998) protocol. DNA from fresh feces that were stored in ethanol and refrigerated was extracted using the commercial QIAamp[®] Fast DNA Stool Mini Kit (QIAGEN, Hilden, Germany) following manufacturer's instructions. All PCRs reactions were carried out in an automatic TC 9639 Thermal Cycler (Benchmark Scientific) in a mixture of final volume of 15 uL containing 3 ng/ul of sample genomic DNA, 7.5 μ L of ImmomixTM mastermix (Bioline), 0.5 μ L of each primer (10 μ M, Table 1) and ultra-pure water.

The profile consisted of an initial denaturation at 95° C for 10 min, followed by 35 cycles of denaturation at 95° C for 1 min, an annealing step for 2 min (Table 1), an extension at 72°C for 1.5 min, and a final extension at 72°C for 7 min. Positive and negative controls were included in every PCR to check for contamination in different experiments.

PCR products were purified with DNA Clean and Concentrator[™] (Zymo Research[™]) kit and diluted to an equal final concentration using a Nanodrop 1000[™] Spectrophotometer (Thermo Fisher Scientific). The amplicons were sequenced by the Sanger method on an automatic sequencer ABI 3130 (Applied Biosystems) at the Pasteur Institute (Montevideo, Uruguay) and on an automated ABI 3730xl System Sequencer (MACROGEN Inc., Korea).

Bioinformatic Analysis. Sequences were aligned and edited in MEGA11 (Molecular Evolutionary Genetics Analysis software version 11,<u>Tamura *et al.* 2021</u>) and compared with the nucleotide database available in the National Cen-

ter for Biotechnology Information (https://www.ncbi.nlm. nih.gov/) using BLAST (Basic Local Alignment Search Tool) utility. We calculated the number of paired differences for each mitochondrial marker. Diversity indexes as haplotype, nucleotide diversities and the number of polymorphic sites were calculated using DnaSP v. 6.12.3 (Rozas *et al.* 201714). The evolutionary genetic distances among haplotypes were determined by the Kimura 2-Parameter distance (Kimura 1980). Nucleotide sequence data were analyzed using maximum parsimony (MP) and neighbor joining (NJ) in the software MEGA11 (Tamura *et al.* 2021), as well the cluster support trees were evaluated with 1000 bootstrap pseudo-replicates.

Haplotype networks. To evaluate the evolutionary relationships among haplotypes and their geographic distribution, we constructed two haplotype networks (from *Cytb* and *D-loop* sequences) using the median-joining network approach (<u>Bandelt *et al.* 1999</u>) implemented on PopART 1.7 (Leigh and Bryant 2015).

Patterns of geographic subdivision and gene flow. We used AMOVA (Analysis of Molecular Variance) to deduce the significance of geographic divisions among local and regional population groupings (Excoffier et al. 1992; Schneider et al. 2000). We calculated the fixation index within populations (Φ ST) and among populations (Φ SC), as well as within groups and between groups (Φ CT). We also estimated the average number of migrants per generation (using Φ ST estimates between populations) to measure the degree of isolation of populations or the degree of subdivision among populations. The significance of F-statistic analogues was evaluated by 1,000 random permutations of sequences among populations. We experimented with various grouping of populations as suggested by the analysis of DNA sequences and population trees. The groupings which maximized values of Φ CT and were significantly different from random distributions of individuals were assumed to be the most probable geographical subdivisions (Excoffier et al. 1992; Schneider et al. 2000).

Gene flow within and among regions was approximated as Nm, the number of female migrants occurring between populations units per generation and was estimated using the expression Fst = 1/ (1+2Nm) where N is the female effective population size and m is the female migration rate (Slatkin 1987; 1993; Baker *et al.* 1994). We used pairwise estimates of Φ ST as surrogates for FST among regional groupings of populations (*e. g.* Stanly *et al.* 1996).



Figure 1. Distribution map of Pampas deer populations. Grey shadow indicated the presumed past distribution of the species. Yellow shaded areas represent the current distribution. Numbers indicated the 8 localities sampled: 1. Rocha; 2. Salto; 3. Samborombón Bay; 4. Santa Fé; 5. San Luis; 6. Emas National Park; 7. Pantanal da Nhecolandia; and 8. Paraná. Asterisk (*) shows ECFA (Estación de Cría y Fauna Autóctona) locales in Maldonado, Uruguay.

Phylogenetics and evolutionary rate estimation of three mtDNA regions. The phylogenetic relationships and evolutionary rates based on three markers (*D-loop*, *Cytb* and *COI*) were estimated using the Neighbour Joining algorithm. Trees were drawn to scale with the length of the branches in the same units as the evolutionary distance used to infer the phylogenetic tree. The genetic distances were calculated using the Kimura 2-Parameters algorithm and the units are the number of base substitutions per site.

We used the divergence time between the Pampas deer *O. bezoarticus* and the grey brocket deer *Subulo gouazoubira* based on <u>Duarte *et al.* (2008)</u>, to estimate the evolutionary rate of the three mitochondrial markers *Cytb*, *COI*, and *D-loop*.

Table 1. Primers sequences annealing temperature and amplicon length.

Mitochondrial region	Primer sequence	Annealing temperature (°C)	Amplicon length (bp)	Reference
COI	LCO1490 5' -GGTCAACAAATCATAAAGATATTGG-3´ HC02198 5' - TAAACTTCAGGGTGACCAAAAAATCA-3´	55	710	Folmer <i>et al.</i> 1994
Cytb	H15149 5'-GCCCCTCAGAATGATATTTGTCCTCA-3' L14724 5'- CGAAGCTTGATATGAAAAACCATCGTTG- 3'	57	480	Maldonado <i>et al</i> . 1995
D-loop	Thr-L 15926 5' -CAATTCCCCGGTCTTGTGAACC -3' DL-H 16340 5' -CCTGAAGTAGGAACCAGATG -3'	50	603	Kocher <i>et al</i> . 1989

The evolutionary rate of *Cytb* was estimated using sequences deposited in GenBank by <u>Duarte et al.</u> (2008; accession numbers: DQ789173-DQ789231) together with additional sequences retrieved from GenBank and employed by the authors in the phylogenetic analysis. To calculate the evolutionary rate of the *D-loop* mitochondrial region, we used Pampas deer sequences and grey brocket deer available in GenBank (Accession numbers: AF012556-AF012602).

Results

We sequenced 138 individuals of the mitochondrial control region (an amplicon of 603 bp). We analyzed a 423 bp fragment and found 87 different haplotypes in the eight localities defined by base-pair substitutions (Table 2). As was reported previously, this species has a polymorphic dinucleotide TA repeat sequence within this amplicon that had four to eight tandem repeats beginning at nucleotide position 186, with position 1 as the first nucleotide of our control region sequence (González *et al.* 1998). Because the same allele sizes were found in divergent sequences from geographically distant populations (*e. g.* Emas and Rocha) this region has a high degree of homoplasy and we excluded the tandem repeat region from the analysis, leaving 423 base pairs of DNA sequence to be analyzed.

Patterns of geographical subdivision and gene flow revealed by D-loop region. We analyzed samples from nine geographic locations that showed high levels of polymorphism (Table 3). The highest value obtained for the average number of pairwise differences within population (PiX) was in the Paraná samples and the lowest in Samborombón Bay.

Geographic distribution of control region sequences. The control region haplotypes were perfectly segregated as no locality shared haplotypes except for the Salto population that shared haplotypes with the individuals from the Breeding center ECFA. Sequences from the same locality tend to be clustered together in minimum spanning networks (Figure 2). The nine groups according to the haplotype's geographic distribution and the taxonomic criteria were arrange in six populations (Groups: 1 Salto and EFCA, 2 Rocha, 3: Paraná. 4: Emas, 5: Pantanal da Nhecolandia and North Argentina, and 6: Samborombón Bay and San Luis (Table 4). This is because the haplotype diversity present in each population, where most of the haplotypes found for each, are unique and not shared with other populations.

The AMOVA results showed high levels of differentiation among populations within groups (Φ ST = 0.439), among groups (Φ SC = 0.240), and as well as within populations (Φ CT = 0.262; Table 5). The Argentinean populations (Samborombón Bay and San Luis) clearly are more closely related to each other than are those from Brazil or Uruguay.

Genetic differences among and within populations. The mtDNA control region showed a rapidly evolving pattern compared with more conserved genes such as the COI and Cytb. The COI gene showed phylogenetic relationships among Pampas deer populations in the recent past, where many migrants per generation were exchanged between populations, and population genetic structure

Table 2. Pampas deer individuals from locations analyzed *Dloop* haplotype, repetitions subspecies and Accession Numbers.

Location	Haplotype	Sequences	Repeated	Subspecies	Acc. number
Rocha- Uruguay	SGO2		1	O. b. uruguayensis	AF012589.1
33°45'S; 54°02'W	SGO7	SG13	2	O. b. uruguayensis	AF012591.1
	SG19	SG15	2	O. b. uruguayensis	AF012590.1
	SG34		1	O. b. uruguayensis	AF012588.1
	SG91		1	O. b. uruguayensis	OR805768
	SG109		1	O. b. uruguayensis	OR528919
	SG118		1	O. b. uruguayensis	OR528922
	SG126		1	O. b. uruguayensis	OR528923
	SG134	SG117, SG123, SG124, SG125, SG127, SG128, SG129, SG130, SG131, SG132, SG133	12	O. b. uruguayensis	OR805767
	SG1738		1	O. b. uruguayensis	AF012597.1
	SG245		1	O. b. uruguayensis	OR528924
	SG329		1	O. b. uruguayensis	OR528927
EL TAPADO- Uruguay	SG01		1	O. b. arerunguaensis	AF012601.1
31°65'S; 56°43'W	SG04	SG60, SG110, SG111, SG251, SG259, SG274, SG289ECFA, SG319, SG320, SG370, SG371, SG372, SG375, SG382, SG383, SG384	17	O. b. arerunguaensis	AF012583.1
	SG09		1	O. b. arerunguaensis	AF012598.1
	SG10		1	O. b. arerunguaensis	AF012586.1
	SG11		1	O. b. arerunguaensis	AF012587.1
	SG16		1	O. b. arerunguaensis	AF012585.1
	SG17		1	O. b. arerunguaensis	AF012602.1
	SG20		1	O. b. arerunguaensis	AF012600.1
	SG49		1	O. b. arerunguaensis	AF012596.1
	SG76	SG328, SG330, SG331, SG377, SG380	6	O. b. arerunguaensis	OR528934
	SG95		1	O. b. arerunguaensis	OR528938
	SG112		1	O. b. arerunguaensis	OR528920
	SG113		1	O. b. arerunguaensis	OR528921
	SG378	SG381	2	O. b. arerunguaensis	OR528929
	SG379		1	O. b. arerunguaensis	OR528930

Table 2. Continuation...

Location	Hanlotype	Sequences	Repeated	Subspecies	Acc number
	sc1622	Sequences	1		ACC: 1101110E1
	561625		1	O b arerunguaensis	OP529025
	50252		1	O. b. arerunguaensis	OR528925
54 46 5; 55 14 W	50281		1	O. D. arerunguaensis	08526920
	56369		1	O. b. arerunguaensis	OR528928
Paraguay	SG385		1	O. b. arerunguaensis	OR528931
Unknown location	SG94		1	O. b. leucogaster	OR528937
PARANA- Brasil	FB39	FB40, FB42, FB49, FB52	5	0. b. sp.	OR528912
24°11'S; 49°46'W	FB38	FB41	2	0. b. sp.	OR528911
	FB46	FB47, FB48	3	O. b. sp.	OR528913
	FB51		1	O. b. sp.	OR528914
EMAS- Brasil	SP13	SP12	2	O. b. bezoarticus	AF012558.1
18°15'S; 52°53'W	SP14	SP18	2	O. b. bezoarticus	AF012559.1
	SP15		1	O. b. bezoarticus	AF012560.1
	SP17	SP20	2	O. b. bezoarticus	AF012561.1
	SP19		1	O. b. bezoarticus	AF012599.1
	SP51		1	O. b. bezoarticus	AF012592.1
	SP52		1	O. b. bezoarticus	AF012562.1
	SP53		1	O. b. bezoarticus	AF012593.1
	SP54		1	O. b. bezoarticus	AF012563.1
	SP55		1	O. b. bezoarticus	AF012564.1
	SP56		1	O. b. bezoarticus	AF012565.1
PANTANAL- Brasil	SP36		1	O. b. leucogaster	AF012566.1
18°15'S: 52°53'W	SP38		1	O. b. leucogaster	AF012567.1
,	SP41		1	0 h leucogaster	AF012568 1
	SP42		1	0 h leucogaster	AF012569.1
	SP/3		1	0. b. leucogaster	AF012571 1
	5045		1	0. b. leucogaster	AF012570.1
	5040	CD4E CD47 CD40 CD40	5	O. b. leucoguster	AF012570.1
	SP40	5845, 5847, 5846, 5849	5	O. b. leucoguster	AF012572.1
	5P40		1	O. b. leucoguster	AF012575.1
SAMBOROMBON-	SPSU		1	O. D. leucogaster	AF012574.1
Argentina	SG24		1	O. b. celer	AF012581.1
35°30'S; 56°45'W	SG39		1	O. b. celer	AF012594.1
	SG40		1	O. b. celer	AF012578.1
	SG42		1	O. b. celer	AF012579.1
	SG43		1	O. b. celer	AF012580.1
	SG52		1	O. b. celer	AF012582.1
	SG72		1	O. b. celer	OR528933
SAN LUIS- Argentina	SG18	SG84, SG105	3	O. b. celer	AF012576.1
34°22'S; 65°44'W	SG66	SG85, SG88	3	O. b. celer	AF012575.1
	SG67		1	O. b. celer	AF012595.1
	SG68		1	O. b. celer	AF012577.1
	SG83		1	O. b. celer	OR528935
	SG86		1	O. b. celer	OR528936
	SG104		1	O. b. celer	OR805766
	VEN006		1	O. b. celer	OR528939
	VEN012		1	O. b. celer	OR528940
	VEN013		1	O. b. celer	OR528941
	VEN015		1	O. b. celer	OR528942
	VEN035		1	O. b. celer	OR528943
	VEN040		1	O. b. celer	OR528944
	VEN049		1	O. b. celer	OR528945
	VEN057		1	O. b. celer	OR528946
	VEN061		1	O. b. celer	OR528947
	VEN067		1	O. b. celer	OR528948
	VEN070		1	0. b. celer	OR528949
SANTA FE- Argenting	P\/2		1	0 h leurogaster	OR528916
31°32'C+ 60°41'M/	D\/2		1	0. b. leucogaster	OR528017
51 50 5, 00 41 W	D\/7		1	0. b. leucogaster	OR538019
CORRIENTES- Argenting	PV/11		1	0. b. leurogaster	OR528915

 Table 3.
 Above diagonal: Average number of pairwise differences between populations (PiXY).
 Diagonal elements: Average number of pairwise differences within population (PiX).
 Below diagonal: Corrected average pairwise difference (PiXY-(PiX + PiY)/2).

	1	2	3	4	5	6	7	8	9
1	6.009	6.872	10.710	8.014	13.920	10.893	8.975	11.619	15.208
2	2.261	3.213	10.824	7.445	12.923	11.303	9.234	8.312	11.598
3	5.666	7.178	4.079	9.093	14.633	11.552	10.923	14.953	18.434
4	3.792	4.620	5.836	2.435	10.627	9.139	9.159	11.706	15.481
5	5.489	5.889	7.167	3.982	10.853	14.031	12.354	16.087	20.244
6	3.778	5.586	5.402	3.810	4.493	8.222	11.593	15.540	18.929
7	1.593	3.250	4.506	3.563	2.549	3.104	8.756	12.437	15.844
8	4.816	2.907	9.116	6.691	6.863	7.631	4.261	7.596	12.433
9	5.299	3.087	9.489	7.359	7.913	7.913	4.561	1.730	13.810

1. Rocha; 2. Salto; 3. ECFA; 4. Samborombóm Bay; 5. North Argentina (Santa Fe); 6. San Luis; 7. Emas National Park; 8. Pantanal da Nhecolandia; and 9. Paraná.

was almost non-existent. On the other hand, the characteristic *D-loop* hypervariable region reflects the phylogenetic relationships among contemporary populations of Pampas deer, where populations are genetically isolated without gene flow.

The results of the different mitochondrial markers showed that the *D-loop* has unique haplotypes within populations. The haplotype diversity value for *COI* gene was 0.35, the average nucleotide diversity by site was $\pi = 0.00391$ (s. d. = 0.0000003). The populations that exhibited the highest values were Paraná, Rocha, San Luis, and Salto.

We found 18 haplotypes of 63 samples in the *Cytb* gene sequence fragment of 417 bp in Pampas deer populations, and a diversity index of 0.28 (Table 6). These three mito-chondrial markers provide resolution at different scales and allow us to elucidate different scenarios and the spatial connections of haplotypes to make inferences about demographic and evolutionary processes of this species (Figure 2, Table 7).

On the other hand, we found lower evolutionary rates for the *COI* gene than *Cytb*. From the nucleotide differences matrix between pairs of *COI* sequences we determined 18 haplotypes. The minimum spanning network constructed has the H1 and H5 haplotypes with an ancestral position and has a wide distribution (Figure 2). This is consistent in terms of geographic location of the different populations. The number of migrants per generation that is exchanged between these populations is less than one migrant per generation, in the case of Paraná and Santa Fe was the lowest value of migrants, Nm = 0.286.

The AMOVA results for the *CO*I region analysis showed a lack of structure amongst populations. To explain this, we put forward the hypothesis that this lack of structure detected among populations is that millions of years ago these populations were connected. The mitochondrial control region (*D-loop*) analysis shows a later stage to this connection between haplotypes of the populations occurring after isolation by habitat fragmentation. The existence of high genetic variation featuring Pampas deer today indicates that the decrease in population size was recent. **Table 4.** Migrants estimations (M = Nm for haploid data) in bold above diagonal. The average number of pairwise differences within population (PiX) is shown on the diagonal numbers in italic. Below diagonal, Slatkin (1995) linearized FSTs as t/M=FST/(1-FST).

	1	2	3	4	5	6
1	6.14159	0.48088	0.76385	0.66852	1.41493	0.96551
2	1.03976	4.07854	0.31033	0.44385	0.70567	0.46762
3	0.65458	1.61120	2.43542	0.92304	1.18.19	0.74569
4	0.74792	1.12650	0.54169	10.85335	1.54293	0.83842
5	0.35337	0.70854	0.42116	0.32406	9.68892	0.98286
6	0.51786	106.924	0.67052	0.59636	0.50872	12.79634

References: Groups 1. El Tapado and EFCA, 2. Los Ajos, 3. Paraná, 4. Emas, 5. Pantanal and North Argentina, and 6. Buenos Aires and San Luis Provinces.

Discussion

Past Genetic Variation. The direct ancestor of the Pampas deer first appeared in the Pampean Formation during the Pleistocene and may be associated with a glacial event approximately 2.5 million years ago at the boundary between the Gauss and Matuyama chrons (Bonadonna and Alberdi 1987; Marshall *et al.* 1983). For analyzing the past genetic variation, we used two mitochondrial markers the *Cytb* and *COI* genes.

The divergence time estimated between the two species, Pampas deer and gray brocket deer (*Subulo gouazoubira*) was approximately 4.77 million years. The evolutionary rate found for the *Cytb* gene is 1.07 x10⁻⁸ than it found



Figure 2. Minimum Spanning haplotype networks of mitochondrial regions showing the genetic relationships between Pampas deer individuals: A) Cytb b) COI and C) D loop. Mutational steps between haplotypes are shown as marks across connection lines. Circle sizes are proportional to haplotype frequencies and colours represent the respective sample localities.

Table 5. Summary of Analysis of Molecular Variance (AMOVA) results for *D-loop* region analysis in different groupings of the populations. Fixation indexes: between groups (Φ SC), among populations within groups (Φ ST), within populations (Φ CT).

D.f.	Sum of squares	Variance components	Percentage of variation	Fixation indexes
5	201 277	1 60008 \/2	26.25	ΦSC : 0.240022
5	501.577	1.09008 Va	20.25	<i>p-value</i> = 0.00000
3	38.823	1.14089 Vb	17.72	ΦST : 0.43964 <i>p-value</i> = 0.00000
126	454 650	2 (00 40)/-	56.04	ФСТ: 0.2624
120	454.058	3.60840 VC	50.04	<i>p-value</i> = 0.00000
134	794.858	6.43937		
	D.f. 5 3 126 134	D.f. Sum of squares 5 301.377 3 38.823 126 454.658 134 794.858	D.f. Sum of squares Variance components 5 301.377 1.69008 Va 3 38.823 1.14089 Vb 126 454.658 3.60840 Vc 134 794.858 6.43937	D.f. Sum of squares Variance components Percentage of variation 5 301.377 1.69008 Va 26.25 3 38.823 1.14089 Vb 17.72 126 454.658 3.60840 Vc 56.04 134 794.858 6.43937 3.000000000000000000000000000000000000

Note: Distance method: Kimura 2P. p < 0.000001, after 1023 permutations.

for the *COI* gen 9.6 x10 ⁻⁹ nucleotide substitution events per site per year would be happening (See Figure 2). The sequences for the *D* loop regions analyzed with 423 base pairs fragments showed an evolutionary rate of 1.4 x10 ⁻⁸. The obtained divergence values for Pampas deer using the two mitochondrial markers, show that, as expected, the *D*-loop evolves faster than the *COI* gene.

The inference through the calculated evolutionary rate is that the *COI* gene could reflect another stage of the Pampas deer population history being more appropriate to analyze phylogenetics relationships. *Cytb* and *COI* genes are not able to resolve recent demographic events linked to habitat fragmentation (Tobe *et al.* 2010). Estimation of migrants per generation showed a high gene flow among the populations studied, and because of this, we found less genetic population structure according to the *COI* gene.

On the other hand, nucleotide diversity per site for Pampas deer through the COI gene is 0.391 %. The Paraná population had the highest nucleotide diversity value (0.87 %) while the lowest one was evidenced in the Emas population (0.16%). The nucleotide diversity among populations of Pampas deer, through analysis of the control region obtained by González et al. (1998), ranged between 1.1 % to 2.5 %, and Argentina's population had the lowest value. Their slower evolutionary rate than D-loop showed that existing populations in the past maintained a closer relationship, showing that the structure is a recent phenomenon and could be a consequence of genetic populations' isolation. A similar finding was observed in another endangered species like the Asian elephant (Elephas maximus), whose populations are rapidly declining and *D-loop* marker was suitable for analyzing genetic variation and to infer other processes such as introgression/hybridization (Srikulnath et al. 2023).

Our results confirm the previous findings by <u>González et</u> <u>al. (1998)</u> that the control region of the Pampas deer is one the most polymorphic of any mammal. The large number of haplotypes and high level of nucleotide diversity in the Pampas deer suggest that this species was more abundant and widespread in the recent past. Since variability was lost rapidly from these populations, sizes have remained small

Haplotype	Individuals	Repetitions	Subspecies	Accession Number
H1	SG38, SG52,	15	O. b. arerunguaensis	OR659038
	SG95, SG100, SG95, SG100, SG105, SG123, SG144, SG147,		O. b. bezoarticus	
			O. b. celer,	
	SG150, SG207, SG217, PV11,		O. b. leucogaster	
	FB03		O. b. uruguayensis	
			O. b. sp.	
H2	SG85, SG103	2	O. b. celer	OR659039
H3	SG83, SG99, SG102	3	O. b. celer	OR659040
H4	SG104	1	O. b. celer	OR659041
H5	PV2, PV3, PV7	12	O. b. arerunguaensis	OR659042
	SP13, SP15,		O. b. bezoarticus	
	SP31, SP32, SP56, SG227,		O. b. leucogaster	
	SG48, SG49, SG1623		O. b. uruguayensis	
H6	SP35, SP36, SP37, SP46, SP50	5	O. b. leucogaster	OR659043
H7	SP48	1	O. b. leucogaster	OR659044
H8	SP40	1	O. b. leucogaster	OR659045
H9	FB52	1	O. b. sp.	OR659046
H10	FB47, FB49	2	O. b. sp.	OR659047
H11	FB51	1	O. b. sp.	OR659048
H12	SG143	1	O. b. uruguayensis	OR659049
H13	SG13	1	O. b. uruguayensis	OR659050
H14	SG109	1	O. b. uruguayensis	OR659051
H15	SG10	1	O. b. arerunguaensis	OR659052 OR659053
H16	SP27	1	O. b. bezoarticus	
H17	SP14	1	O. b. bezoarticus	OR659054
H18	SP25	1	O. b. bezoarticus	OR659055
TOTAL		51		

References: Haplotype (H) identification, individuals ID, subspecies and the GenBank accession numbers.

Table 6. Pampas deer haplotypes of partial COI gen.

Table 7. Pampas deer haplotypes of partial Cyt b gen.

Haplotype	Sequences	Repetitions	Subspecies	Acc. number
			O. b. arerunguaensis	
			O. b. bezoarticus	
I	SG02, SG04, SG07, SG13, SG18, SG40, SG43, SG60, SG105, SG111, SG112, SG113, SG126, SG1623, SG320, SP19, P15, PV3, PV6	19	O. b. celer	MH593537.1
			O. b. leucogaster	
			O. b. uruguayensis	
П	SG09	1	O. b. arerunguaensis	OR546559
Ш	SG19, SG34, SG91, SG117, SG123, SG124, SG125, SG127, SG128, SG129, SG130, SG131, SG132, SG133, SG134, SG1738	16	O. b. uruguayensis	OR546557
IV	SG44	1	O. b. arerunguaensis	OR546556
V	SG94	1	O. b. leucogaster	OR546555
VI	SG95, SG378	2	O. b. arerunguaensis	OR546554
VII	SG109	1	O. b. uruguayensis	OR546558
VIII	SP22, SP53	2	O. b. bezoarticus	OR546553
IX	SP36, SP42	2	O. b. leucogaster	OR546552
х	SP38	1	O. b. leucogaster	OR546551
XI	SP55	1	O. b. bezoarticus	OR546550
XII	FB21, FB38, FB41, FB48, FB49, FB52	6	O. b.sp	OR546563
XIII	FB39	1	O. b. sp	OR546562
XIV	FB51	1	O. b. sp	OR546561
XV	PV2	1	O. b. leucogaster	OR546560
XVI	VEN035	1	O. b. celer	OR546549
XVII	VEN012	1	O. b. celer	DQ789191.2
XVIII	VEN057VEN013, VEN015, VEN049, VEN040	5	O. b. celer	OR546548
Total		63		

References: Left column number of haplotype identification, the repetitions, subspecies and the GenBank accession numbers. Total number of Cyt b sequences obtained.

for long periods of time (Ballou 1994). However, currently Pampas deer are endangered in Argentina, south of Brazil and Uruguay, with fewer than 2,500 mature individuals. The levels of genetic diversity in populations from these locations suggest that historic population sizes were several orders of magnitude larger, and that recently populations have decreased dramatically, thus providing a strong mandate for restoration and augmentation. This population decline was due to habitat loss and unregulated hunting beginning in the last century and, most recently, to control efforts by ranchers who believe that deer compete with livestock. Pampas deer numbers might increase if protected from poaching in areas where natural habitats remain and if some grazing land, as a buffer, could be designated for dual use by deer and livestock (Castro *et al.* 2021).

González *et al* (1998) had estimated the historic population size based on the relation $\boldsymbol{\Theta} = 2N\mu$ where N is the effective number of females and μ is the mutation rate per site. Using coalescent likelihood methods incorporated in the COALESCE program by Kuhner *et al.* (1995), the parameter $\boldsymbol{\Theta}$ can be calculated from a population sample of DNA sequences. Our estimate of $2N\mu$ is 0.173 and assuming a mutation rate of 2.5 x 10⁻⁸ per nucleotide site per year for the control region (based on sequence divergence between the Pampas and brocket deer, as above), the effective number of breeding females would be approximately 3,460,000. The total census size of females is likely to be at least double this value (*e. g.* Nunney and Elam 1994). Therefore, both comparative and theoretical estimates indicate a substantial reduction in population size has occurred since the total

present-day number of deer is between 64,000 and 80,000 individuals (González et al. 2010).

Conservation Implications. Genetic units for conservation have been based on criteria such as reciprocal monophyly (evolutionary significant units) or differences in genotype frequency (management units; <u>Moritz 1995</u>). The numerous reticulations in the minimum spanning network (Figure 2) show that none of the neighboring Pampas deer populations are reciprocally monophyletic and indicate the occurrence of past episodes of migration.

The highest level of genetic differentiation was observed in Uruguayan populations, especially the Rocha population. This differentiation is explained, with the UPGMA algorithm, establishing a divergence time of 2 million years and considering mean genetic distance and *D-loop* mutation rate (Figure 3). Probably this separation is linked to events occurring during the Pleistocene.

However, except for the Argentinean populations from Samborombón Bay and San Luis, belonging to *O. b. celer*, and North Argentinean and Pantanal belonging to *O. b. leucogaster*, all the other populations are significantly or marginally differentiated, thus they might be classified as management units experiencing low to modest rates of gene flow. A pronounced sequence divergence exists between Brazilian populations from Emas National Park and Pantanal da Nhecolandia and North Argentinian populations (Figure. 3), corresponding to the different subspecific designation recognized by <u>Cabrera (1943</u>). These populations may have been historically isolated in different habitats. In fact, the population in Emas National Park is located in the



Figure 3. Evolutionary relationships of populations. The optimal tree with the sum of branch length = 1.43701786 is shown. The tree is drawn to scale, with branch lengths in the same units as those of the evolutionary distances used to infer the phylogenetic tree. The evolutionary distances were estimated with Kimura 2 parameter and the analyses were conducted in MEGA11.

cerrado of central Brazil, 650 to 1,000 m elevation, which has a distinct dry season, whereas the population in El Pantanal is found in wetlands below 100 m. Although based on limited evidence, these physiological differences may indicate differences in the timing of the reproductive cycle and hence, if of genetic origin, may be an important reason why the populations should not be interbred or used as a source for cross-translocation. Supporting differentiation between these populations are discrepancies in their physiology. In the Emas National Park population, antlers are shed in April, whereas in the Pantanal this occurs in June and July (González et al. 2010). The individuals belonging to Paraná showed differentiated genetic distance and low number of migrants less than 1. This population is critically endangered, making it urgent to reduce activities such as poaching that may be severely affecting its conservation, and to promote conservation management actions and design mitigation measures to assure long-term survival.

Our results suggest that Pampas deer have the potential to exist over a much greater area and historical data demonstrate a much wider distribution for the species. Therefore, if the goal of conservation is to maintain long term population stability and preserve genetic variation, conservation efforts should focus on the restoration of deer habitats over a wide geographic area. Finally, we conclude that the genetic dynamic shown by Pampas deer allows us to identify the *D-loop* as the marker of choice for defining management units for conservation of this species.

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Literature cited

- AVISE, J. C. 1992. Molecular population structure and the biogeographic history of a regional fauna: a case history with lessons for conservation biology. Oikos 63:62-76.
- Avise, J. C. 1995. Mitochondrial DNA polymorphism and a connection between genetics and demography of relevance to conservation. Conservation Biology 9:686-690.
- Avise, J. C. 1998. The history and purview of phylogeography: a personal reflection. Molecular Ecology 7:371-379.
- BALLOU, J. D. 1994. Population biology. Pp. 93–114, *in* Przewalski 's horse: the history and biology of an endangered species (Boyd, L., and K. A. Houpt, eds.). State University of New York Press. New York, U.S.A.
- BAKER S.C., W. R. SLADE, and J. L. BANNISTER. 1994. Hierarchical structure of mitochondrial DNA gene flow among humpback whales, Megaptera novaeanglie, world wide. Molecular Ecology 3:313-327.
- BANDELT, H. J., P. FORSTER, and A. RÖHL. 1999. Median-joining networks for inferring intraspecific phylogenies. Molecular Biology Evolution 16:37-48.
- BONADONNA, F. P., and M. T. ALBERDI, M. T. 1987. *Equus stenonis* Cocchi as a biostratigraphical marker in the Neogene-Quaternary of the Western Mediterranean basin: consequence on Galerian-Villafranchian chronostratigraphy. Quaternary Science Reviews 6:55-66.
- CABRERA, A. 1943. Sobre la sistemática del venado y su variación individual y geográfica. Revista del Museo de La Plata. (n.s) Tomo III Zoología 18: 5-41.
- CASTRO, M., *ET AL*. 2021. Parasites of Pampas deer (*Ozotoceros bezoarticus* L. 1758) in sympatry with livestock in Uruguayan Agroecosystems. Revista Argentina de Parasitología10:7–14.
- DARWIN, C. 1860. A naturalist's voyage round the world. The voyage of the Beagle. Natural history library edition, 1962. Doubleday and Company, Inc., Garden City, New York., U.S.A.
- DUARTE, J. M. B., S. GONZÁLEZ, and J. E. MALDONADO. 2008. The surprising evolutionary history of South American deer. Molecular Phylogenetics and Evolution 49:17-22.
- Excoffier, L., P. Smouse, and J. QUATTRO. 1992. Analysis of molecular variants inferred from metric distances among DNA haplotypes: Application to human mitochondrial DNA restriction data. Genetics 131: 479-491.
- GIMÉNEZ DIXON, M. 1986. Ozotoceros bezoarticus (L.,1758) In: Identification manual: Mammalia (Carnivora to Artyodactyla) (Dollinger, P., ed.) Compiled with the advice and guidance of Manual Comitee) 1:119.006.007.001.
- GONZÁLEZ, S., ET AL. 1998. Conservation genetics of the endangered Pampas deer (*Ozotoceros bezoarticus*). Molecular Ecology 7:47-56.

- GONZÁLEZ, S., *ET AL*. 2002. Morphometric differentiation of endangered Pampas deer (*Ozotoceros bezoarticus*), with description of new subspecies from Uruguay. Journal of Mammalogy 83:1127-1140.
- GONZÁLEZ, S., *ET AL.* 2016. Conservation genetics, taxonomy, and management applications in neotropical deer. Tropical Conservation: Perspectives on Local and Global Priorities 1:238-254.
- GONZÁLEZ, S., *et al.* 2023. Supporting the recovery of the Pampas deer in Uruguay. Oryx 57:423-424.
- GONZÁLEZ, S., ET AL. 2010. Pampas deer Ozotoceros bezoarticus (Linnaeus 1758). Pp. 119–132, *in* Neotropical Cervidology: biology and medicine of Latin American deer. Jaboticabal, Brazil.
- GONZÁLEZ, S., J. J. Jackson, and M. L. MERINO. 2016. *Ozotoceros bezoarticus*. The IUCN Red List of Threatened Species: e. T15803A22160030. Accessed on 09 January 2024.
- JACKSON, J. E., and A. LANGGUTH. 1987. Ecology and status of the Pampas deer in the Argentinean Pampas and Uruguay. Pp. 402-409, *in* Biology and Management of the Cervidae En (Wemmer C. M., ed.). Research symposia of the National Zoological Park Smithsonian Institution. Front Royal, U.S.A.
- JACKSON, J. E. 1987. *Ozotoceros bezoarticus*. Mammalian Species 295:1-5.
- KIMURA, M. 1980. A simple method for estimating evolutionary rates of base substitutions through comparative studies of nucleotide sequences. Journal of Molecular Evolution 16:111-120.
- KUHNER, M. K., J. YAMATO, and J. FELSENSTEIN, J. 1995. Estimating effective population size and mutation rate from sequence data using Metropolis-Hastings sampling. Genetics 140: 1421-1430.
- LEIGH, J. W., and D. BRYANT. 2015. POPART: full-feature software for haplotype network construction. Methods in Ecology and evolution 6:1110-1116.
- Marshall, L. G., R. HOFFSTETTER, and R. PASCUAL. 1983. Mammals and stratigraphy: geochronology of the continental mammalbearing Tertiary of South America. Palaeovertebrata 13: 1-93.
- MARQUEZ, A., ET AL. 2006. Phylogeography and Pleistocene demographic history of the endangered marsh deer (*Blastocerus dichotomus*) from the Río de la Plata Basin. Conservation Genetics 7:563-575.
- MERINO, M. L., *ET AL*. 1997. Veado campeiro (*Ozotocerus bezoarticus* Linneus 1758). In Biología e Conservacao de Cervideos Sul-Americanos; *Blastocerus*, *Ozotoceros* e *Mazama*. Jaboticabal, FUNEP. Jaboticabal, Brasil.
- MORITZ, C. 1995. Uses of molecular phylogenies for conservation. Philosophical Transactions of the Royal Society of London. Series B: Biological Sciences 349:113-118.
- NUNNEY, L., and D. R. ELAM. 1994. Estimating the effective population size of conserved populations. Conservation Biology 8:175-184.
- PINDER, L. 1994. Status of Pampas deer in Brazil. Population and Habitat Viability Pp. 157-162, *in* Assessment for the Pampas deer (*Ozotoceros bezoarticus*),(González S, M. Merino, M. Gimenez-Dixon, and E. S., Seal). Workshop Report CBSG/ IUCN. Apple Valley, U.S.A.
- ROZAS, J., *ET AL*. 2017. DnaSP 6: DNA sequence polymorphism analysis of large data sets. Molecular Biology and Evolution 34:3299-3302.

- SCHNEIDER, S. D. ROESSLI, and L. EXCOFFIER. 2000. Arlequin version 2.000: a software for population genetics data analysis. Suiza: University of Geneva. Geneve, Switzerland.
- SLATKIN, M. 1987. Gene Flow and the geographic structure of natural populations. Science 236:787-792.
- SLATKIN, M. 1993. Isolation by distance in equilibrium and non-equilibrium populations. Evolution 47:264-279.
- SLATKIN, M. 1995. A measure of population subdivision based on microsatellite allele frequencies. Genetics 139:457-62.
- SNEATH, P. H., and R. R. SOKAL. 1973. Numerical taxonomy. The principles and practice of numerical classification. Freeman and Co. San Francisco, U.S.A
- STANLY, H. F., *ET AL*. 1996. Worldwide patterns of mitochondrial DNA differentiation in the harbord seal (*Phoca vitulina*). Molecular Biology and Evolution 13:368-382.
- SRIKULNATH, K., *ET AL*. 2023. Asian Elephant Evolutionary Relationships: New Perspectives from Mitochondrial D-Loop Haplotype Diversity. Sustainability 15:720.
- TABERLET, P., *ET AL*. 1996. Reliable genotyping of samples with very low DNA quantities using PCR. Nucleic Acids Research 24:3189-3194.
- TAMURA, K., G. STECHER, and S. KUMAR. 2021. MEGA11: molecular evolutionary genetics analysis version 11. Molecular Biology and Evolution 38:3022-3027.
- TOBE, S. S., *ET AL*. 2010. Reconstructing Mammalian Phylogenies: A Detailed Comparison of the Cytochrome b and Cytochrome Oxidase Subunit I Mitochondrial Genes. Plos One 5:e14156.
- WEBER, M., and S. GONZALEZ. 2003. Latin American deer diversity and conservation: a review of status and distribution. Ecoscience 10:443-454.

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Supplementary material

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